
Extracellular bacterial production of doped magnetite nanoparticles

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DOI: 10.1039/9781849734844-00102

Microorganisms have been producing nanoparticles for billions of years and by controlling and tuning this productivity they have the potential to provide novel materials using environmentally friendly manufacturing pathways. Metal-reducing bacteria are a particularly fertile source of nanoparticles and their reduction of Fe (III) oxides leads to the formation of ferrite spinel nanoparticles, especially magnetite, Fe₃O₄. The high yields produced by extracellular biomineralising processes make them commercially attractive, and the production of these bionano ferrite spinels can be tuned by doping the precursor Fe(III) phase with Co, Ni, Zn, Mn and V. The oxidation state of the cations and the sites of substitution are determined by X-ray absorption spectroscopy (XAS), especially by examination of metal *L*-edge spectra and X-ray magnetic circular dichroism (XMCD). Vanadium substitution in bionano ferrite spinels is revealed for the first time, and substitution in the octahedral site as V(III) confirmed. Bionanomagnetite is shown to be effective in the remediation of azo dyes with the complete breakdown of Remazol Black B to colourless amines and acids. XMCD shows this to involve oxidation of the surface Fe(III) and the potential for regeneration of the nanoparticles.

1 Introduction

The current search for functional nanomaterials is unrelenting with a wide variety of inorganic pathways being employed and investigated. However, it is becoming clear that microbial approaches, exploiting a wide and relatively untapped genetic diversity that has developed over > 3.5 billion years of evolution, offers potentially clean and scalable biotechnological alternatives.¹ Physiological processes including redox transformations linked to microbial respiration of metals, detoxification reactions, and the bioproduction of biological ligands that can precipitate metals in highly reactive compartments in or around the microbial cell, can all play their part in nanoparticle biosynthesis reactions. For a recent overview of some of the products produced by such processes, which include bionanoparticles developed for catalysis, antimicrobial, photonic and magnetic applications, the reader is referred to^{2,3} and references therein. This review focuses on the microbial production of iron-based magnetic nanoparticles, which can be

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produced either internally as “magnetosomes” which form as an integral part of the cell in a very controlled metabolic environment, or as a result of extracellular process where the production of nanoparticles is a by-product of respiratory processes, in this case the “dissimilatory” reduction of ferric iron minerals.^{4–6} Although first recognised in the proteobacteria, the ability of microorganisms to produce intracellular Fe-based magnetotactic nanoparticles may well have evolved several times in the evolutionary record,⁵ and examples of extracellular production of nanoparticles by dissimilatory iron reduction occur in many archaeal and bacterial genera, catalysed by distinct mechanisms that are genera specific.^{6–8}

The intracellular magnetosomes found in magnetotactic bacteria which occupy a range of aquatic environments, have been studied extensively. The nanoparticles classically form chains of regular single crystals of magnetite (Fe_3O_4) or less regular crystals of greigite (Fe_3S_4).^{9,10} These are contained within the cell and the 30–150 μm magnetite particles form monodisperse, perfect cubic {100}, octahedral {111} and dodecahedral {110} forms, or hybrids of these shapes such as cubooctahedral habits^{11,12} and therefore present the potential for the production of high quality nanocrystals (Fig. 1).

Furthermore, a relationship between bacterial species and crystal shape, and the successful doping of the magnetite with 1.4% Co^{13,14} gives the prospect of customised nanoparticles. However, despite their biological and intrinsic interest, the problem of exploiting magnetosomes is the fastidious nature of these fascinating (but often hard to cultivate) organisms, and the challenge of producing enough biomass to produce even milligrams of material.

In terms of volume, it is the extracellular production of nanoparticles by bacteria that would seem to have most commercial potential.¹⁵ Over the past decades, our knowledge of extracellular biomineralising processes and microorganisms has grown in response to our appreciation of their role in elemental cycling (discussed at length in^{16,17}). Passive interactions of cells with their geochemical environment causes biomineralisation either by metal sorption onto charged cell walls and extracellular layers, resulting in heterogeneous nucleation and mineral precipitation, or by microbial metabolism changing the local redox conditions that results in the expulsion of ligands from the cell¹⁷ (see Fig. 4 below). Extracellular nanoparticles are

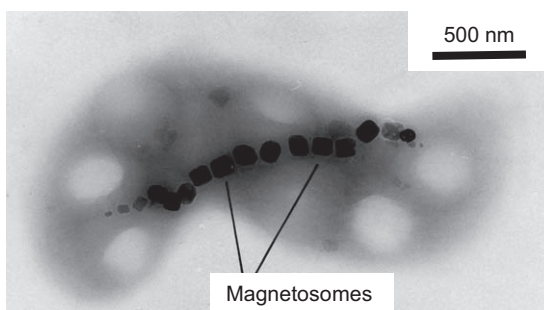


Fig. 1 Magnetosomes in a magnetotactic *Spirillum species*. The single magnetic domain, magnetite crystals form a chain across the bacterium and have typical elongate cubooctahedral {100} + {111} habit.

produced by bacteria in a range of natural environments and isolates have been examined under a number of experimental conditions.

A range of oxides, phosphates, carbonates and elemental nanoparticles can be produced. Of special interest to nanotechnologists are the enzymatically controlled redox changes that result in biomineral formation which are linked to microbial respiration such as dissimilatory metal reduction.³ This latter process has been shown, for instance, to produce Ag(0) nanoparticles 5–40 μm in size^{19,20} (Fig. 2), selenium/selenide/telluride nanospheres and rods^{21,22} (Figs. 3 and 4), Au(0)²³, Pd(0)²⁴ as well as Tc(IV) and U(IV)²⁵ (see also²⁶ for recent reviews).

The enzymatic reduction of soluble U(VI) to produce insoluble U(IV) nanoparticles and Tc(VII) to produce stable Tc(IV) are also important bioremediation pathways^{27–29} with potential for the clean-up of legacy contaminated soils, and relevance to the geodisposal of radwastes.

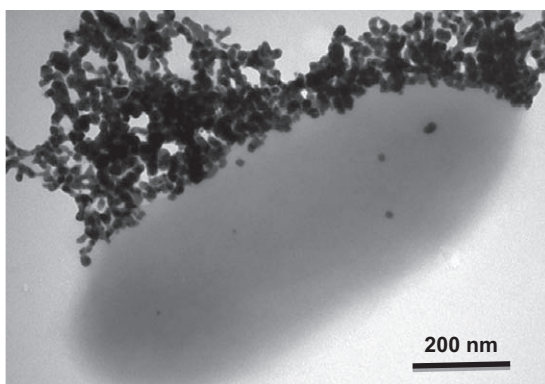


Fig. 2 Ag(0) nanoparticles produced by enzymatic reduction of Ag(1)Cl by *Geobacter sulfurreducens* (adapted from Ref. 18).

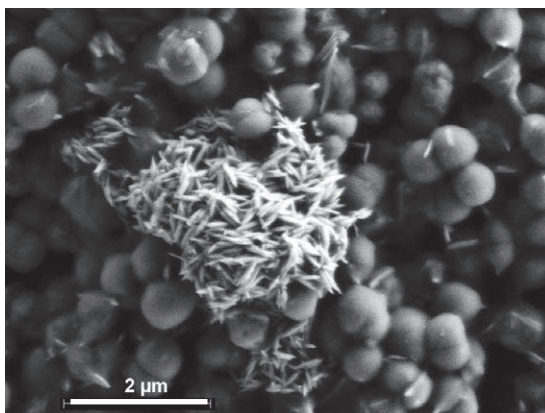


Fig. 3 Te-nanorods. *Veillonella atypica* cannot use Te-oxyanions as a terminal electron acceptor for anaerobic respiration but forms ~ 100 nm Te(0) nanorods, which can be seen protruding from the spherical cells, and as extracellular clusters, when grown under anaerobic conditions in the presence of Te(IV) at 37 °C in a medium containing yeast extract and lactate.

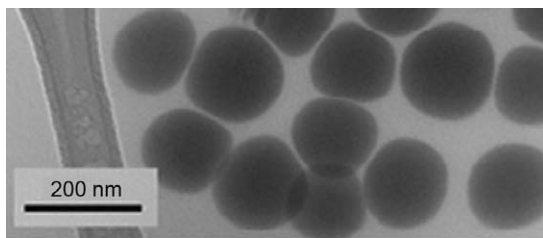


Fig. 4 Selenium nanoparticles produced by the reduction of sodium selenite by *Geobacter sulfurreducens*³² which can also reduce Se(IV) and Te(IV), with c-type cytochromes implicated in the electron transfer to the metalloid.

The challenge is to harness these bio-manufacturing capabilities and produce nanoparticles with tuneable properties of high enough quality and large enough quantity to challenge the environmentally more damaging, current methods of synthesis. The panacea would be the fabrication of exploitable particles from deleterious wastes, resulting in remediation with a saleable by-product; examples are the production of saleable ferrite spinels from Fe-ferrihydroxides derived from acid mine drainage or precious metals extracted from electronic wastes.¹ A further challenge is to control the chemistry of the nanoparticles. The addition of metals other than Fe into the structure of nanomagnetite (Fe_3O_4) is well known to enhance greatly the magnetic properties of the particles, making them suitable for a range of industrial applications.^{30,31}

2 Exploiting extracellular biogenic magnetite

Magnetite nanoparticles are of particular interest because of their technological value^{33–37} and they are produced by many anaerobic Fe(III)-reducing bacteria found in the subsurface, where they respire the Fe(III) phase, coupling metal reduction to the oxidation of naturally occurring organic compounds or hydrogen to conserve energy for growth.^{5,27,38} The most intensively studied Fe(III)-reducing bacteria include *Geobacter* and *Shewanella* species, although there are many other phylogenetically distinct microorganisms able to respire Fe(III) minerals³⁹ forming nanosized magnetite (Fe_3O_4) as a stable end-product (Fig. 5).⁴⁰ Here we summarise existing work and present a new study of the substitution of V into bionanomagnetite and a demonstration of the effectiveness of biomagnetite in technological processes, in this case the remediation of azo dyes. The value of X-ray absorption spectroscopy (XAS) in nano-materials characterisation is also demonstrated, especially soft X-ray analysis of the transition metal *L*-edges and X-ray magnetic circular dichroism (XMCD).

3 Metal doped magnetites

Over the past 5 years doped extracellular bionanomagnetite has been produced successfully in several model bacterial systems.^{41–44} First row transition metals Co, Ni, Mn, Cr and Zn have all been successfully incorporated into the bionanomagnetite structure using Fe(III)-reducing bacteria

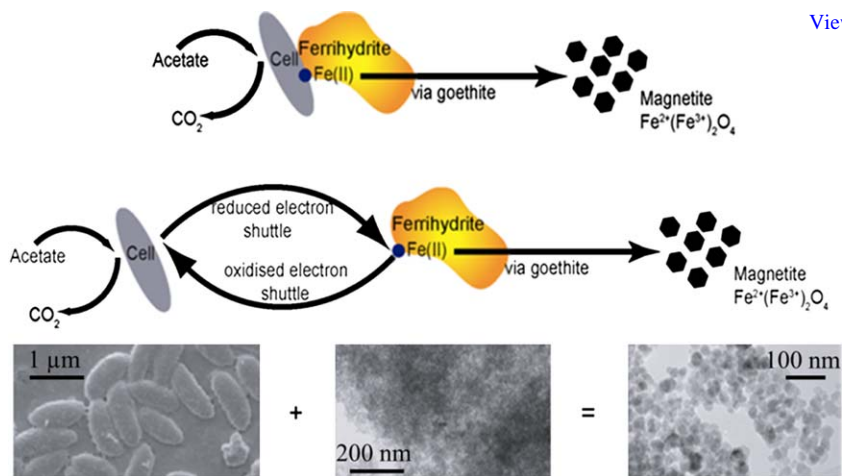


Fig. 5 The production of magnetite by the bacterium *G. sulfurreducens*. Cells oxidize organic substrates such as acetate to produce CO_2 and directly donate electrons to the extracellular Fe(III) phase, such as schwertmannite or ferrihydrite but this requires close contact between the cell surface and the mineral phase. The rate of reduction can be increased by addition of the humic analogue anthraquinone-2,6-disulfonic acid (AQDS), which is reduced by the cells of *G. sulfurreducens* and acts as an electron shuttle between the cell surface and the Fe(III)-bearing mineral. The AQDS donates electrons to the Fe-hydroxyoxide, reducing Fe^{3+} to Fe^{2+} . Subsequent reaction between the Fe^{2+} and the remaining Fe^{3+} results in the formation of a biogenic mixed $\text{Fe}^{2+}/\text{Fe}^{3+}$ -bearing mineral phase (top), magnetite (Fe_3O_4) (middle). The transmission electron micrographs reveal that the particle size of the biogenic magnetite is ~ 30 nm (bottom).^{40,44}

including *G. sulfurreducens*, *Shewanella species* and the thermophile *Thermoanaerobacter ethanolicus* TOR-39. The precursor solid-phase used has been Fe(III)-bearing minerals such as akaganeite ($\text{Fe}^{3+}\text{O}(\text{OH},\text{Cl})$) or ferrihydrite ($\text{Fe}^{3+}_2\text{O}_3 \cdot 0.5\text{H}_2\text{O}$), containing the appropriate dopant, as the electron acceptor for microbial Fe(III)-reduction^{41–54} (see Table 1 for details).

During bioreduction, up to one-third of the Fe in Fe_3O_4 has been substituted by first row transition metals resulting in significant changes in the magnetic properties of the resulting biogenic ferrite spinels.^{46,49} For example, *G. sulfurreducens* produced Co-ferrite with a magnetic coercivity at 5 K of 7900 Oe, compared to the equivalent ‘unsubstituted’ biogenic magnetite value of only 360 Oe; values equivalent to the chemically synthesized counterparts.⁴⁶ Other elements have also been bio-incorporated into the spinel structure to a lesser but still significant degree. The rare earth metals including Nd, Gd, Tb, Ho and Er (= R) can be included in the magnetite spinel structure using *T. ethanolicus* TOR-39 and *S. putrefaciens* PV-4 to give particles with the chemical formula up to $\text{R}_{0.06}\text{Fe}_{2.94}\text{O}_4$.^{15,49,50} The contaminant arsenate was also found to be substituted into the magnetite spinel structure up to 1% of the total cations using *G. sulfurreducens*, which has implications for arsenic mobility in the subsurface.⁵¹ It should be noted that Co, Ni, Mn and Zn have also been incorporated into intracellularly produced magnetite by magnetotactic bacteria, however the substitution levels were much lower than those achievable using extracellular formation processes.^{14,55,56}

Table 1 Production of doped bionanomagnetites.

| Dopant | Bacterium | Reference |
|----------------|--|--|
| Cobalt | <i>Geobacter sulfurreducens</i> | Coker <i>et al.</i> , 2008 ⁴⁵ ; Coker <i>et al.</i> , 2009 ⁴⁶ |
| | <i>Shewanella oneidensis</i> MR-1 | Coker <i>et al.</i> , 2008 ⁴⁵ |
| | <i>Shewanella putrefaciens</i> PV-4 | Moon <i>et al.</i> , 2007a ⁴³ ; Moon <i>et al.</i> , 2007c ⁴⁹ |
| | <i>Thermoanaerobacter ethanolicus</i> TOR-39 | Moon <i>et al.</i> , 2010b ⁵⁰ ; Roh <i>et al.</i> , 2001 ⁵² ; Roh <i>et al.</i> , 2006 ⁵² ; Moon <i>et al.</i> , 2007a ⁴³ ; Moon <i>et al.</i> , 2007c ⁴⁹ |
| Nickel | <i>Geobacter sulfurreducens</i> | Coker <i>et al.</i> , 2008 ⁴⁵ |
| | <i>Shewanella oneidensis</i> MR-1 | Coker <i>et al.</i> , 2008 ⁴⁵ |
| | <i>Shewanella putrefaciens</i> CN32 | Fredrickson <i>et al.</i> , 2001 ⁴⁷ |
| | <i>Shewanella putrefaciens</i> PV-4 | Moon <i>et al.</i> , 2007a ⁴³ ; Moon <i>et al.</i> , 2007c ⁴⁹ |
| | <i>Thermoanaerobacter ethanolicus</i> TOR-39 | Moon <i>et al.</i> , 2010b ⁵⁰ ; Roh <i>et al.</i> , 2001 ⁵² ; Roh <i>et al.</i> , 2006 ⁵¹ ; Moon <i>et al.</i> , 2007a ⁴³ ; Moon <i>et al.</i> , 2007c ⁴⁹ |
| Chromium | <i>Shewanella putrefaciens</i> PV-4 | Moon <i>et al.</i> , 2007a ⁴³ ; Moon <i>et al.</i> , 2007c ⁴⁹ |
| | <i>Thermoanaerobacter ethanolicus</i> TOR-39 | Roh <i>et al.</i> , 2001 ⁵² ; Roh <i>et al.</i> , 2006 ⁵¹ ; Moon <i>et al.</i> , 2007c ⁴⁹ |
| Manganese | <i>Geobacter sulfurreducens</i> | Coker <i>et al.</i> , 2008 ⁴⁵ |
| | <i>Shewanella putrefaciens</i> PV-4 | Moon <i>et al.</i> , 2007a ⁴³ ; Moon <i>et al.</i> , 2007c ⁴⁹ |
| | <i>Thermoanaerobacter ethanolicus</i> TOR-39 | Roh <i>et al.</i> , 2006 ⁵¹ ; Moon <i>et al.</i> , 2007a ⁴³ ; Moon <i>et al.</i> , 2007c ⁴⁹ |
| Zinc | <i>Geobacter sulfurreducens</i> | Coker <i>et al.</i> , 2008 ⁴⁵ |
| | <i>Shewanella putrefaciens</i> PV-4 | Moon <i>et al.</i> , 2010a ¹⁵ ; Moon <i>et al.</i> , 2007a ⁴³ ; Moon <i>et al.</i> , 2007c ⁴⁹ |
| | <i>Thermoanaerobacter ethanolicus</i> TOR-39 | Moon <i>et al.</i> , 2007a ⁴³ ; Moon <i>et al.</i> , 2007c ⁴⁹ ; Moon <i>et al.</i> , 2010a ¹⁵ ; Yeary <i>et al.</i> , 2011 ⁵⁴ |
| Manganese/Zinc | <i>Geobacter sulfurreducens</i> | Coker <i>et al.</i> , 2008 ⁴⁵ |
| Arsenic | <i>Geobacter sulfurreducens</i> | Coker <i>et al.</i> , 2006 ⁵³ |
| Vanadium | <i>Geobacter sulfurreducens</i> | This paper |
| Rare Earths | <i>Thermoanaerobacter ethanolicus</i> TOR-39 | Moon <i>et al.</i> , 2010b ⁵⁰ ; Moon <i>et al.</i> , 2007a ⁴³ ; Moon <i>et al.</i> , 2007c ⁴⁹ |

4 X-ray magnetic circular dichroism (XMCD)

Crucial to the understanding of metal-substituted magnetite and their properties is the determination of the site occupancy of the metals. The structure of magnetite is very well known. It is a cubic, inverse spinel with one quarter of the tetrahedral (T_d) and one half of the octahedral (O_h) sites filled by Fe. The formula for magnetite is $(Fe^{3+})_4[Fe^{2+}Fe^{3+}]O_4$ where Fe^{3+}

is equally split between (T_d) and [O_h] sites. Fe^{2+} occupies only O_h sites resulting in a distribution for stoichiometric magnetite of 1:1:1 for the Fe^{2+} O_h : Fe^{3+} T_d : Fe^{3+} O_h sites. The Fe^{3+} cations in the (T_d) and [O_h] sublattices have antiparallel magnetic moments that compensate each other, but there is a resulting net ferromagnetic effect due to the Fe^{2+} on the O_h sites. The synchrotron radiation technique, X-ray magnetic circular dichroism (XMCD) has proved a powerful tool in determining the site occupancies of ferrites such as magnetite. XMCD spectra are derived from the difference between absorption spectra collected for right and left circularly polarized light, or opposite applied magnetic fields.⁵⁷ It is sensitive to the oxidation state and local structure of magnetically ordered iron cations at solid surfaces.⁵⁸ In magnetite the XMCD shows three distinct sharp features at the Fe L_3 absorption edge which, from low to high photon energy, correspond to the Fe^{2+} O_h , Fe^{3+} T_d , and Fe^{3+} O_h site contributions, respectively.⁵⁸ The Fe^{2+} and Fe^{3+} ions at the O_h sites are aligned ferromagnetically, while coupled antiferromagnetically to the Fe^{3+} at the T_d sites, which displays the opposite sign in the XMCD (see Fig. 6b). The relative site occupancies of Fe in the three sites and the effect of additional metals is then determined by quantifying the peak intensities and comparing with standard samples as well as through fitting using multiplet calculations,^{58,59} providing detailed information on the magnetite structure and site occupancy.

5 Vanadium biomagnetite

Vanadium substitution on spinel ferrites is of technological interest because it can be used to tailor the magnetic and electrical properties of the nanoparticles. It is found in nature associated with titanium-bearing magnetites and is an important source of vanadium for ferroalloys. Vanadium is toxic to higher life forms⁶⁰ and anthropogenic contamination can occur *via* the petroleum industries because it is a trace component of fuel oils,⁶¹ and is present in certain uranium ores, thus is concentrated in former processing sites.⁶¹ Incorporation of vanadium in bionanomagnetites has not been previously examined. Many metal reducing bacteria are able to couple the reduction of vanadium(V) to the oxidation of organic matter.^{5,62,63} In this study, a hydrous ferric oxide containing V(V) was synthesised by coprecipitation and then V-substituted biomagnetite produced using *G. sulfurreducens* coupled to the oxidation of sodium acetate using the methods described in.⁴⁴ The magnetic precipitate was characterised using transmission electron microscopy (TEM) and XMCD in order to determine the quantity, valence state and site occupancy of the Fe and V. Figure 6(a) reveals the particle size of the magnetite to be ~ 25 nm. XAS spectra monitored in total-electron yield (TEY) mode were collected on beamline I06 at Diamond Light Source using the portable octopole magnet system endstation.⁶⁴ At each energy point the XAS was recorded for the two opposite magnetisation directions by reversing the applied field of 0.6 T. The XAS spectra of the two magnetisation directions were normalised to the incident beam intensity and the XMCD is obtained as the difference spectrum.⁵⁸ The Fe and V $L_{2,3}$ XMCD gives the site occupancies and

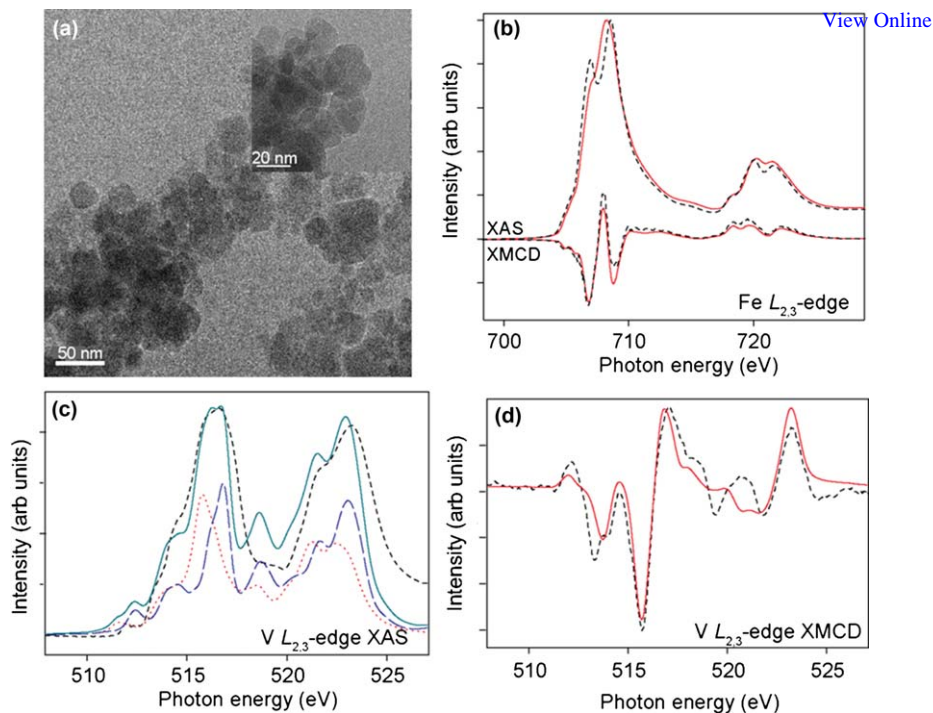


Fig. 6 Vanadium substitution in bionanomagnetite. a) Transmission electron micrographs of 10% vanadium doped magnetite; b) Fe $L_{2,3}$ XAS and corresponding XMCD spectra for biomagnetite (solid line (red)) and 10% vanadium-doped magnetite (dashed); c) calculated V $L_{2,3}$ -edge XAS for V(III) (dots (red)) and V(IV) (long dashes (blue)), and V(III) + V(IV) summed in a 1:1 ratio (solid line (green)) compared with the experimental XAS (dashes (black)); d) calculated XMCD for V(III) (solid line (red)) compared with experiment (dashed (black)). The V spectra were calculated for octahedral crystal field of $10Dq = 1.8$ eV in the absence of $3d$ spin-orbit interaction.

oxidation state of the cations in the spinel structure of the magnetite, while the relative XAS intensities give the concentration ratio of the two elements. To obtain the relative amounts of Fe in the three coordination sites, the experimental XMCD spectra were fitted by means of a non-linear least-squares analysis, using the calculated spectra for each of the different Fe sites (see^{57–59} for details).

Using the integrated intensities from the V and Fe $L_{2,3}$ XAS, the quantity of vanadium contained in the V-biogenic magnetite is 8.9 at% compared to iron. The shape of the Fe L_3 XAS, when compared to that for biomagnetite without vanadium (Fig. 6b), suggests that a portion of the ‘non-magnetic’ precursor material remains unconverted to magnetite and is represented by the prominent peak at the low energy side of the main Fe L_3 peak. The Fe $L_{2,3}$ XMCD of the biogenic magnetite and V-substituted biogenic magnetite was fitted to obtain the Fe^{2+}/Fe^{3+} ratios as 0.61 and 0.88, respectively. This result indicates that vanadium is substituting for Fe^{3+} . The T_d/O_h ratios for biomagnetite and V-biomagnetite are 0.44 and 0.55, respectively, indicating that the Fe^{3+} substitution is occurring predominantly in octahedral sites. Examination of the vanadium $L_{2,3}$ XMCD and XAS spectra provides additional information. The V $L_{2,3}$ XMCD (Fig. 6d) shows

an excellent agreement with the calculated spectrum for V(III), and since V_2O_3 does not contribute to the XMCD as it is antiferromagnetic (critical temperature = 160 K), the presence of the XMCD confirms that V(III) is incorporated in the spinel structure. Comparison between the measured and calculated V $L_{2,3}$ XAS^{65,66} indicates that, since the peaks at 516.5 eV and 523 eV are split, the vanadium is present in the sample both as V(III) and V(IV). The calculated XAS spectra, where the V(IV) peak has 1–2 eV higher energy than the V(III) peak, produce a good fit for V(III):V(IV) = 1:1 (Fig. 6c) indicating that, in addition to V(III) incorporated in the ferrite spinel, a VO_2 component is present. Previous work using Fe K -edge XAS on natural V-bearing magnetite⁶⁷ containing a few wt% V, suggests the vanadium to be largely present as V(III) with less than 10% V present as V(IV). This study shows that the bacterial reduction of the doped precursor produces vanadium bionanomagnetite nanoparticles containing 8.9 wt% V and the combined Fe and V XMCD and XAS shows that it is present as V(III) replacing Fe(III) predominantly in the octahedral sites. As V is present in soils as the bioavailable and toxic V(V) or in more reduced soils as V(IV), this incorporation in magnetite, and as V(III) provides a potential immobilisation and remediation pathway for contaminated sites.

6 Bionanomagnetite in textile wastewater treatment

The potential of bionanomagnetite in remediation of radionuclide and chromium contamination has been demonstrated^{3,27–29} but less has been undertaken on organic pollutants. Here this is addressed by the testing of the performance of bionanomagnetite to remediate azo dyes. The chromophore in reactive azo dyes consists of azo/keto-hydrazone groups as in the model reactive azo dye, Remazol Black B. Low levels of dye–fibre fixation (Fig. 7), and the presence of unreactive hydrolysed dye in the dyebath, lead to losses of up to 50% of the dye to the wastewater. These problems are compounded by the high water solubility and characteristic brightness of the dyes. Due to their stability and to their xenobiotic nature, reactive azo dyes are not totally degraded by conventional wastewater treatment processes that involve light, chemicals or activated sludge.^{68–70} The dyes are therefore released into the environment in the form of coloured wastewater. In this study, nanoscale schwertmannite $[Fe^{3+}_{16}O_{16}(OH)_{12}(SO_4)_2]$ powder was synthesized as a model Fe(III)-oxide starting material.⁷⁰ The schwertmannite was reduced under anaerobic conditions using *G. sulfurreducens* (see Fig. 5), in the presence of acetate as the electron donor. The resulting bionanomagnetite phase was isolated, washed and, along with commercially available ‘abiotic’ nanoscale magnetite (Johnson Matthey), exposed to Remazol Black B.

The changes in the dye were monitored by measuring the optical density at the λ_{max} for the dye (597 nm) using a UV-visible spectrophotometer. The effect of the biogenic nanomagnetite on the Remazol Black B was dramatic, revealed by the change in shape and intensity of absorption spectrum, with a major reduction in absorption over 4 days and the production of a colourless solution in 31 days that was stable to oxidation (Fig. 8).

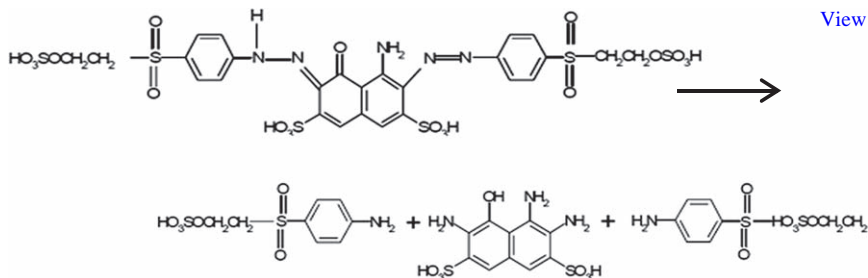


Fig. 7 Structure of the dark blue/black reactive azo dye Remazol Black B. The azo/keto-hydrazone groups can be reduced to produce the corresponding colourless amines Para Base and Diamino H-acid.

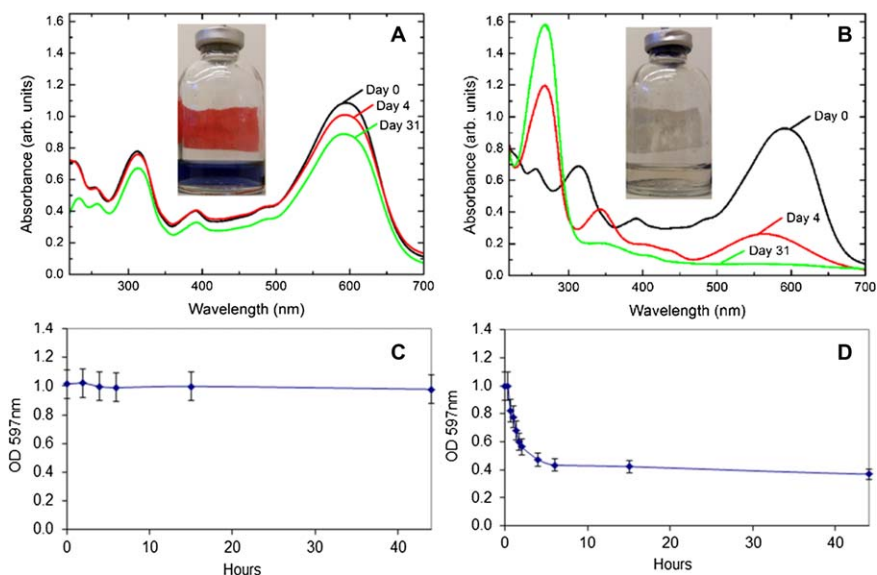


Fig. 8 Interaction of bionanomagnetite with azo dyes. UV-visible absorption spectra and inset showing colour of solution after reaction for inorganic nanomagnetite (A) and biogenic nanomagnetite (B) showing the fast and effective breakdown of the azo dye in the latter. The rate of Remazol Black B reduction by inorganic nanomagnetite (C) and biogenic nanomagnetite (D) is revealed graphically by monitoring the absorption at 597 nm.

The bionanomagnetite had totally reduced the azo chromophore in the dye to form the corresponding colourless amines. Commercially available abiotic magnetite adsorbed the dye, resulting in a slight reduction in intensity but no change in shape of the absorption spectrum and the solution remained distinctly blue.

Both bionanomagnetite and abiotic nanomagnetite were analysed, before and after dye reduction using XMCD (Fig. 9) and show that the $\text{Fe}^{2+}/\text{Fe}^{3+}$ ratio at the surface of the abiotic nanomagnetite did not change during the experiment, but the biogenic nanomagnetite showed a decrease in the amount of Fe^{2+} , indicating the Fe^{2+} was oxidized concurrent with the reduction of Remazol Black B. However, a substantial

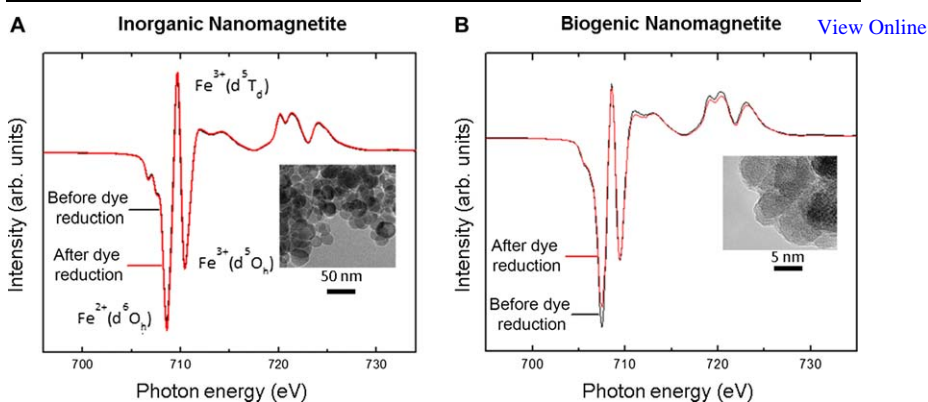


Fig. 9 XMCD spectra of bionanomagnetite before and after interaction with the azo dye. (A) shows the XMCD spectra (inset showing TEM image of the nanoparticles) for inorganic magnetite and (B) for biogenic nanomagnetite. The latter shows reduction of the octahedral Fe^{2+} component, as indicated (black curves are before interaction, grey (red) curves are after interaction).

amount of Fe^{2+} remained at the surface of the biogenic nanomagnetite after the reductive transformation, presenting the opportunity for recycling the material.

7 Conclusions

This review outlines some recent advances in our understanding of the bioproduction of functional nanoparticles. Although focusing on the tunable bioproduction of nano-scale extracellular magnetite, this work should be seen in the context of recent dramatic advances in the microbial fabrication of a far wider range of functional nanomaterials, reviewed in,³ and extended recently to form a new area of synthetic biology.⁷¹ To date much of this work has been done at the laboratory scale, producing mg quantities of material, and successful scale-up is now an important priority. This has been achieved for Zn-substituted biomagnetite using *T. ethanolis* TOR-39 in a fermenter (30 L/1 kg dry weight magnetite),¹⁵ and more recently by this group using both *Shewanella* and *Geobacter* species at a similar scale. Many challenges remain. For those working with intracellular bacterial magnetosomes the particles produced are consistent in size and shape and the challenge is volume of production, while for extracellular bacterial nanoparticles, volumes are already achievable but particle homogeneity and narrower size ranges are the challenge. New metagenomic strategies to identify magnetosome genes with potential templating techniques may lead to high value specialised nanoparticles from magnetotactic bacteria.^{72–75} Production of extracellular nanoparticles will tap into the enormous range and diversity of biomineralising bacterial species and their high productivity, adaptability and tunability; these are likely to produce materials for less specialised applications and bioremediation.

Acknowledgements

The authors are grateful to support from an EPSRC collaborative research fund to support this work. Portions of this work were performed at the

Molecular Foundry and Beamline 4.0.2 at the Advanced Light Source, [View Online](#) part of the Lawrence Berkeley National Laboratory, which is supported by the Office of Science, Office of Basic Energy Sciences, US Department of Energy, under contract no. DE-AC02-05CH11231, with special thanks to Ron Zuckermann (Biological Nanostructures Facility) and Elke Arenholz (ALS). XMCD measurements were also undertaken on beamline I06 at the Diamond Light Source, Didcot, Oxfordshire, UK and the beamline staff are thanked for their generous support. The magnetite for the azo dye experiment was prepared by Richard Cutting, SEAES, University of Manchester, UK.

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